

**Science Approach Document:  
Ecological Risk Classification of Inorganic Substances**

**Environment and Climate Change Canada**

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## Synopsis

Environment and Climate Change Canada (ECCC) has characterized inorganic substances from the third phase of the Chemicals Management Plan (CMP) for their potential to cause ecological harm. The ecological risk classification of inorganic substances (ERC-I) was applied to a broad group of inorganic substances. These substances either met the categorization criteria under subsection 73(1) of the Canadian Environmental Protection Act, 1999 (CEPA) or were considered a priority on the basis of other human health concerns.

This Science Approach Document presents the ERC-I approach and the results of its application only to inorganic substances identified as having a low level of ecological concern. Inorganic substances not identified as being of low ecological concern by the ERC-I approach will be subject to a more refined analysis in other publications. A period of consultation on this Science Approach Document is being provided in advance of the use of this information in screening assessments to allow the public to comment and provide additional information. Publication of the ERC-I approach and its results in this Science Approach Document will assist the government in addressing substances in a more efficient manner.

The ERC-I involved the use of empirical, measured, and modelled data to classify inorganic substances as warranting further evaluation of their potential to cause harm to the environment or as having a low likelihood of causing ecological harm. Preliminary predicted no effect concentrations (PNEC) were identified or derived to characterize hazard. Predictive models using input from multiple sources and water quality monitoring data from federal and provincial programs were used to characterize exposure and identify predicted environmental concentrations (PECs). Risk quotients comparing PECs to PNECs were calculated, and multiple statistical metrics were applied to derive preliminary classifications for both predictive modelling and water quality monitoring. A risk matrix was used to assign a low, moderate or high classification of potential ecological concern for each substance or group on the basis of the outcome of the modelling and monitoring profiles. Inorganic substances classified as having high ecological concern were generally those characterized as having larger and more frequent PECs exceeding the PNECs. Substances classified as being of low ecological concern would generally be associated with smaller magnitude and infrequent, if any, PECs exceeding the PNECs in both modelled and measured datasets.

Considering inherent hazard properties, current use patterns and quantities in commerce, as well as the analysis of water quality monitoring datasets, 80 substances were classified as being of low ecological concern. Substances that were classified as being of low ecological concern primarily on the basis of current low exposures may be subject to follow-up or tracking of use pattern information to inform future priority-setting.

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# 1. Introduction

Following categorization of substances on the Domestic Substances List (DSL), which was completed in 2006, approximately 4 300 of the 23 000 substances on the DSL were identified for additional assessment activity. These substances met the categorization criteria for persistence or bioaccumulation and inherent toxicity to human or non-human organisms or for greatest potential for exposure to humans under subsection 73(1) of Canadian Environmental Protection Act, 1999 (CEPA) (Canada 1999) or were identified as having health effects of concern based on classifications by other national or international agencies for carcinogenicity, genotoxicity, developmental toxicity or reproductive toxicity. The ecological risk classification of inorganic substances (ERC-I) approach described in this report was applied broadly to inorganic substances identified as remaining priorities for the third phase of the Chemicals Management Plan (CMP). However, this report presents detailed results only for the 80 substances that were classified by ERC-I as being of low ecological concern (hereafter referred to as “ERC-I substances”). Substances classified as being of moderate or high ecological concern will undergo further assessment of ecological considerations in other publications. Although the classifications apply to individual Chemical Abstracts Service Registry Numbers (CAS RN<sup>1</sup>), many of the 80 substances discussed in this report have been further sub-grouped on the basis of common inorganic moieties<sup>2</sup> of potential concern.

The purpose of this Science Approach Document is to give stakeholders and the public the opportunity to review and comment on the ERC-I approach and the results of its application to the substances classified as being of low ecological concern prior to the use of these results—in conjunction with any other relevant information that becomes available after the publication of the Science Approach Document—as the basis for conclusions of screening assessments to be published under section 68 or 74 of CEPA. The publication of the scientific approach and results in the Science Approach Document will assist the government in addressing substances that may be of low concern to either human health or to the environment in a more efficient manner and identifies substances of relatively higher concern that require more detailed evaluation.

The ERC-I approach includes consideration of information on ecological hazards, uses and exposure. The ERC-I was applied using data collected during DSL categorization, through the DSL Inventory Updates, and from federal and provincial water quality monitoring programs, the National Pollutant Release Inventory (NPRI), and other sources. The approach involved the use of empirical, measured, and modelled data to

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<sup>1</sup> The Chemical Abstracts Service Registry Number (CAS RN) is the property of the American Chemical Society, and any use or redistribution, except as required in supporting regulatory requirements and/or for reports to the Government of Canada when the information and the reports are required by law or administrative policy, is not permitted without the prior written permission of the American Chemical Society.

<sup>2</sup> For the purpose of this document, “moiety” signifies a part of a molecule. A moiety is a discrete chemical entity, identified from a parent compound or its transformation products, that is expected to have toxicological significance.

identify the relative likelihood of substances to cause ecological harm. Preliminary PNECs were compared to PECs from predictive modelling approaches using input from several data sources as well as water quality monitoring data from federal and provincial programs. Risk quotients were calculated, and multiple statistical metrics were applied to derive preliminary classifications for predictive modelling and water quality monitoring. These preliminary classifications were combined to reach the overall ecological risk classification for each substance.

Engineered nanomaterial forms of these substances were not explicitly considered in the exposure scenarios of the ERC-I approach. While measured concentrations in the environment could include engineered nanomaterial forms of these substances, engineered nanomaterial forms may be subject to separate assessment considering their unique properties.

This approach focusses on information critical to determining whether or not a substance is likely to cause ecological harm. For the 80 substances identified by ERC-I as having a low likelihood of causing ecological harm, these results are intended to form the basis for the ecological portion of the screening assessments that will be published in conjunction with the assessment of potential human health risks. The basis of the classification pertaining to some of the substances in ERC-I may be subsequently updated and new data considered as part of future assessments.

This document was prepared by staff in the CEPA Risk Assessment Program at Environment and Climate Change Canada and incorporates input from other programs within the department. The document has undergone external written peer review and consultation. Comments on the technical portions of the document were received from Dr. Peter Campbell (Institut national de la recherche scientifique, INRS), Mr. Geoff Granville (GCGranville Consulting Corp.), Dr. Carrie Rickwood (Natural Resources Canada), and Dr. Kevin Wilkinson (Université de Montréal). While external comments were taken into consideration, the final content and outcome of the report remain the responsibility of Environment and Climate Change Canada.

## **2. Basis of approach to ecological risk classification of inorganic substances**

The ERC-I is a risk-based approach that employs multiple metrics for both hazard and exposure in a weighted consideration of multiple lines of evidence. A simplifying assumption for hazard and exposure characterization in ERC-I is a focus on the aquatic compartment. Hazard characterization was based on a survey of past domestic and international assessment predicted no-effect concentrations (PNECs) and water quality guidelines for inorganic moieties. When no PNEC or water quality guideline was found, hazard endpoint data were collected and, depending on data availability, either a species sensitivity distribution (SSD) or an assessment factor (AF) approach was taken

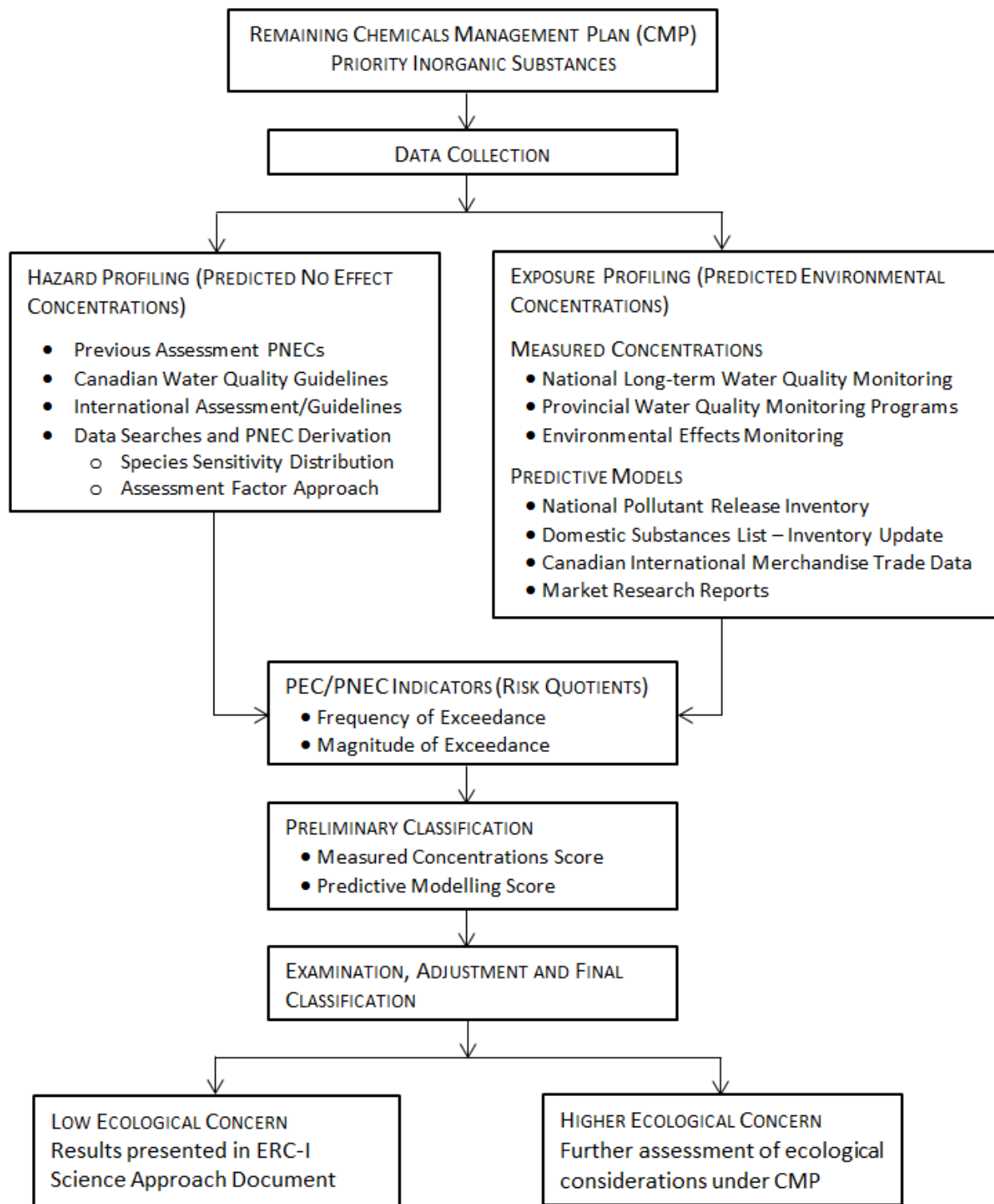
to derive a preliminary PNEC (section 4.1). PNECs for individual remaining priority substances were based on the PNEC identified for the related inorganic moiety.

Exposure profiling considered two approaches: predictive modelling (sections 4.2.2 and 4.2.3) and measured concentrations (section 4.2.4). For predictive modelling, a generic near-field exposure model was used to generate predicted environmental concentrations (PECs) with input data from the NPRI, DSL Inventory Updates, Canada Border Services Agency (CBSA) import data, and third-party market research reports. For measured concentrations, an analysis of measured environmental concentrations for inorganic moieties collected by federal and provincial water quality monitoring programs and the Environmental Effects Monitoring (EEM) Program was conducted. Monitoring datasets for inorganic moieties were used as a conservative estimate of the exposure to individual inorganic substances in the aquatic environment.

Risk quotients (RQ) were calculated by comparing modelled and measured PECs to the PNECs identified during hazard profiling. Multiple statistical metrics were used to classify the potential for ecological risk, considering the magnitude and frequency of observing PNEC exceedances (i.e., RQ greater than 1) and their spatial and temporal distributions. Detection frequency and the natural background ranges (section 4.2.1) were also taken into account in the analysis of measured data. The various lines of evidence were then compared to decision criteria to identify the relative ecological risk of substances and preliminary results for the low ecological concern substances are presented in section 5. This approach reduces the overall uncertainty with risk classification compared to one that relies on a single metric for classification.

The conceptual model for ERC-I is illustrated in Figure 2.1. The preliminary classifications for modelling and monitoring were examined and were adjusted according to specific rules and use of judgement. The final results of ERC-I are then discussed (section 7), along with a summary of uncertainties (section 8).

Critical data and considerations used to create profiles and classifications associated with hazard, exposure and risk are presented in ECCC 2018.



**Figure 2-1. Framework for the ecological risk classification of inorganic substances (ERC-I)**



### 3. Data collection and generation

Substance identity information from past assessment activities was updated with available scientific literature and databases such as National Chemical Inventories (NCI 2015) and SciFinder (2016). For substances with unknown or variable composition, complex reaction products, or biological materials (UVCBs), a chemical structure was chosen to represent the substance. Representative chemical structures to represent an entire UVCB substance (e.g., where variation of the UVCB components was predictable) were selected in a conservative manner by potentially overestimating the percentage weight of the inorganic moiety within the UVCB substance. These percentage weights were subsequently used in predictive modelling calculations.

Ecological hazard information was first collected through a review of recent domestic and international risk assessments. Water quality guidelines, standards and criteria of Canadian federal and provincial governments and other jurisdictions were also considered as ecological hazard data sources where these were available for substances or groups. A comprehensive literature search was conducted for specific substance groups (Ba, Li, Te, Ti, Tl) (Nautilus Environmental 2016), and targeted searches of the ECOTOX database (2016) and European Chemicals Agency (ECHA) registration dossiers (2016) were conducted for substances in the ERC-I approach. In the absence of more recent information, the assumptions used in the 2006 categorization of the DSL were also considered (Canada 2007).

Chemical import and manufacture volumes in Canada and associated use information were collected from the DSL Inventory Update Phase 1 and Phase 2 section 71 surveys for the reporting years 2008 and 2011, respectively (Canada 2009; Canada 2012). For substances where this information was unavailable, trade merchandise data were obtained as direct or surrogate import quantities in Canada. Confidential data to the 10-digit level of the Harmonized System (HS) code were obtained for the years 2010 to 2013 (CBSA 2016). HS 10-digit code descriptions were evaluated for their relationship to ERC-I substances. For example, there is a high correlation between the HS code 2849.20.000, "Silicon carbide," and the remaining priority silicon carbide (CAS RN 409-21-2), and the trade data was taken as direct import quantity for this substance. In another example, data for the HS code 2827.60.1000, "Manganous iodide; potassium iodide; sodium iodide," was taken as a conservative surrogate import quantity, as only potassium iodide (CAS RN 7681-11-0) and sodium iodide (CAS RN 7681-82-5) are remaining priority substances. The Canadian International Merchandise Trade (CIMT) online database, which provides summary data to the HS 6-digit level, was consulted to see if there were significant changes to import quantities in more recent years (CIMT 2016). Where a substance was not surveyed in a DSL Inventory Update, third party market research reports were also obtained as a source of import, manufacture, and use information in Canada (CEH 2014a; CEH 2014b; CEH 2016; MRC 2016).

While import, manufacture, and use information for substances were collected for the purpose of modelling releases to the aquatic environment, reported release data were

also available from the NPRI for certain substances or groups of substances (e.g., lithium carbonate and antimony and its compounds) (NPRI 2016). NPRI data for groups of substances (e.g., antimony and its compounds) were conservatively considered applicable to subsets that are remaining priorities (e.g. 11 antimony-containing substances).

Water quality monitoring data generated by federal and provincial programs covering a number of ecoregions in Canada were collected (Table 3-1). Note that the national long-term water quality monitoring (NLTWQM) datasets cover the Pacific, Northwest Territories, Prairies, and Atlantic regions. As such, data from shared monitoring stations may be duplicated because of federal-provincial monitoring agreements (e.g. Environmental Monitoring System, British Columbia and the NLTWQM Pacific data). Measured concentrations were obtained for the period 2005 to 2015, when available. For some substance groups, measured concentrations in waterbodies exposed to metal mining activities and corresponding reference waterbodies were available from environmental effects monitoring (EEM) studies conducted under the Metals Mining Effluent Regulations (MMER).

**Table 3-1. Summary of federal and provincial water quality monitoring programs and repositories informing the ERC-I approach**

<b>Title</b>	<b>Abbreviation</b>	<b>Reference</b>
Environmental Monitoring System, British Columbia Ministry of the Environment and Climate Change Strategy	EMS	EMSWR 2016
Surface Water Quality Program, Alberta Environment and Parks	SWQP	- <sup>a</sup>
Canada-Alberta Joint Oil Sands Environmental Monitoring	JOSM	JOSM 2016, JOSM 2017
Regional Aquatics Monitoring Program	RAMP	RAMP 2016
Baseline Monitoring of Lower Order Streams in Saskatchewan (BEMLOSS), Saskatchewan Ministry of Agriculture	BEMLOSS	- <sup>b</sup>
Long Term Water Quality Monitoring Network, Government of Manitoba	LTWQMN	- <sup>c</sup>
Provincial Water Quality Monitoring Network, Ontario Ministry of the Environment and Climate Change	PWQMN	PWQMN 2015
Banque de données sur la qualité du milieu aquatique, Government of Quebec	BQMA	BQMA 2015
Surface Water Quality Monitoring, Government of Prince Edward Island	SWQM	- <sup>d</sup>
National Long-term Water Quality Monitoring, Environment and Climate Change Canada	NLTWQM	- <sup>e,f</sup>
Environmental Monitoring and Surveillance	CMPEMS	EC 2009-2012

Program, Chemicals Management Plan		
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<sup>a</sup> Personal communication, data prepared by the Water Policy Branch, Alberta Environment and Parks, for the Ecological Assessment Division, Environment and Climate Change Canada, dated October 2, 2015; unreferenced.

<sup>b</sup> Personal communication, data prepared by the Environmental and Municipal Management Services, Saskatchewan Water Security Agency, for the Ecological Assessment Division, Environment and Climate Change Canada, dated February 25, 2016; unreferenced.

<sup>c</sup> Personal communication, data prepared by the Water Stewardship Division, Province of Manitoba, for the Ecological Assessment Division, Environment and Climate Change Canada, dated February 24, 2016; unreferenced.

<sup>d</sup> Personal communication, data prepared by the Water Quality Monitoring and Surveillance Division, Environment and Climate Change Canada (ECCC) for the Ecological Assessment Division, ECCC, dated September 12, 2016; unreferenced.

<sup>e</sup> Personal communication, data prepared by the Water Quality Monitoring and Surveillance Division, Environment and Climate Change Canada (ECCC) for the Ecological Assessment Division, ECCC, dated September 13, 2016; unreferenced.

<sup>f</sup> Personal communication, data prepared by the Water Quality Monitoring and Surveillance Division, Environment and Climate Change Canada (ECCC) for the Ecological Assessment Division, ECCC, dated February 19, 2016; unreferenced.

## 4. Profiling

### 4.1 Hazard profiling

Profiling of ecological hazard was conducted in a tiered manner for each substance or group of substances. The following sections outline the steps taken to profile ecological hazard. PNEC values for inorganic moieties were read-across to the individual remaining priority substances. Preliminary PNECs for remaining inorganic priorities that did not meet the ERC-I criteria for low ecological concern are not discussed in this Science Approach Document.

#### 4.1.1 Previous risk assessments

For each substance or group of substances, hazard characterizations conducted in screening assessment reports previously published under the Chemicals Management Plan were considered to determine whether PNECs were available and could be applied for ecological risk classification of inorganics. Table 4-1 summarizes the previous hazard characterizations identified as applicable to an ERC-I substance or group of substances.

For antimony-containing substances, the CMP assessment of antimony trioxide was considered for potential hazard alignment (Canada 2010a; EU RAR 2008). The aquatic critical toxicity values (CTV) in the antimony trioxide assessment were based on exposure to antimony trichloride ( $\text{SbCl}_3$ ). Exposure to soluble trivalent antimony (such as  $\text{SbCl}_3$ ) is relevant to both the trivalent and pentavalent remaining antimony-

containing substances. Trivalent antimony largely oxidizes to pentavalent antimony in oxic environments, and there is sparse evidence for a significant difference in ecotoxicological hazard between the two oxidation states (EU RAR 2008; Filella et al. 2009). The aquatic PNEC used in the CMP assessment of antimony trioxide was therefore deemed appropriate for the ecological risk classification of 11 antimony-containing substances (Canada 2010a; EU RAR 2008).

For sodium bromate, the CMP assessment of potassium bromate was identified for hazard alignment (Canada 2010b). The aquatic PNEC in the potassium bromate assessment was derived from consideration of studies conducted almost exclusively on sodium bromate. Both substances readily dissociate to component ions in water. Therefore, the previous hazard characterization was deemed appropriate to apply directly for the ecological risk classification of sodium bromate.

The CMP assessment of vanadium pentoxide was also considered for hazard alignment (Canada 2010c). The species sensitivity distribution constructed in the vanadium pentoxide assessment considered data for several soluble vanadium species: sodium vanadates, vanadium pentoxide, and ammonium vanadate. Therefore, the PNEC was considered appropriate for hazard characterization of two vanadium-containing substances (Canada 2010c).

**Table 4-1. Alignment with previous hazard assessments**

<b>ERC-I substance or group</b>	<b>Previous risk assessment</b>	<b>Previous aquatic PNEC</b>	<b>Uncertainty in application to ERC-I substance or group</b>	<b>Reference</b>
Antimony (11 substances)	Antimony trioxide	113 µg Sb/L	Low. Previous PNEC applicable to both trivalent and pentavalent Sb.	Canada 2010a; EU RAR 2008
Sodium bromate (1 substance)	Potassium bromate	11 µg BrO <sub>3</sub> <sup>-</sup> /L	Low. Sodium bromate was the test substance for the potassium bromate critical toxicity value.	Canada 2010b; Borgmann et al. 2005
Vanadium (2 substances)	Vanadium pentoxide	120 µg V/L	Low. Species sensitivity distribution constructed from multiple pentavalent vanadium species.	Canada 2010c

### 4.1.2 Established water quality guidelines

Searches were carried out to identify established water quality guidelines for substances that could be read-across for ERC-I hazard characterization. Table 4-2 summarizes the established water quality guideline relevant to ERC-I substances classified as low ecological concern.

**Table 4-2. Alignment with established water quality guidelines**

ERC-I substance or group	Established water quality guideline	Guideline value	Uncertainty in application to ERC-I substance or group	Reference
Iron (2 substances)	Iron	300 µg Fe/L	Moderate. Guideline value for total metal in unfiltered sample applied to poorly soluble pigments.	CCREM 1987

### 4.1.3 Deriving new PNECs

Where a previous risk assessment or established water quality guideline was not identified for hazard alignment, new aquatic PNECs were derived. When sufficient data were available, the SSD approach was preferred, following to the extent possible the endpoint selection and data requirement principles of the CCME Protocol for the Derivation of Water Quality Guidelines for the Protection of Aquatic Life (CCME 2007). When insufficient data were available for the SSD approach, PNECs were extrapolated from the lowest acceptable toxicity endpoint (hereafter the critical toxicity value, CTV) using an assessment factor (AF). Assessment factors were determined considering interspecies variability, and normalizations for acute-to-chronic, lethal-to-sublethal, and median-to-low level effects, where required. Derivation details, including the studies identified in literature search activities are further summarized in the ERC-I supporting document (ECCC 2018).

**Table 4-3. Derivation of aquatic PNEC values**

ERC-I substance or group (number of substances)	Approach	Assessment factor	Derived PNEC <sub>aq</sub> (µg/L)
Barium (4)	CTV/AF	5	1780
Beryllium (1)	CTV/AF	10	6.7
Bismuth (7)	CTV/AF	50	0.5
Cerium (1)	CTV/AF	10	3.2
Deuterium oxide (1)	Categorization	1	1000
Germanium (1)	CTV/AF	20	10.5

Hydrogen peroxide (1)	CTV/AF	10	230
Iodine (5)	CTV/AF	100	16.7
Lanthanum (3)	CTV/AF	10	1.8
Lithium (16)	Acute SSD/AF	10	121.6
Molybdenum (2)	Chronic SSD	NA	26340
Neodymium (1)	CTV/AF	50	1.1
Praseodymium (1)	CTV/AF	50	0.7
Silicon carbide (1)	CTV/AF	100	40000
Talc (1)	CTV/AF	100	40000
Tellurium (2)	CTV/AF	10	25
Tin (2)	CTV/AF	5	300
Titanium (13)	CTV/AF	10	850
Yttrium (1)	CTV/AF	5	3.3

NA, not applicable; CTV, critical toxicity value; AF, assessment factor; SSD, species sensitivity distribution.

## 4.2 Exposure profiling

As shown in Figure 2-1, two approaches were used for exposure profiling in the aquatic environment: (1) predictive modelling using a generic near-field exposure model and (2) an analysis of measured concentrations collected by federal and provincial water quality monitoring programs. A weighted approach to exposure profiling was used to address the uncertainty associated with reliance on a single quantitative estimation of chemical release to define exposure to organisms. This helps to mitigate the possibility of overestimation or underestimation of risk classification due to reliance on a single metric (Cimorelli and Stahl 2013). To contribute to this weighted approach, exposure information was collected from multiple sources, as described in section 3. Exposure data availability for the 80 substances in ERC-I are summarized in Appendix A. Predicted environmental concentrations were determined from each source of exposure data as outlined in the subsequent sections.

### 4.2.1 Background ranges

Background ranges were used to establish the environmental context and relevance of the chosen PNECs. Background ranges, or normal ranges of variation, were estimated for inorganic moieties using water quality monitoring samples identified as reference samples by a water chemistry approach predicated on the natural relationship between specific conductivity and alkalinity (Kilgour & Associates Ltd. 2016). The relationship was modelled as a linear regression at the Canadian ecozone and national scale concurrently to classify water quality monitoring samples as being in a reference or non-reference condition. Normal ranges were then derived using reference samples from stations that were always in reference condition (i.e., only containing reference samples) as well as all reference samples within an ecozone. Four types of background ranges for the ecozones resulted: those developed using (1) samples from stations that

were always in a reference condition, as identified by the ecozone-specific conductivity-alkalinity regression; (2) all reference samples, as identified by the ecozone-specific conductivity-alkalinity regression; (3) samples from stations that were always in a reference condition, as identified by the national conductivity-alkalinity regression; (4) all reference samples, as identified by the national conductivity-alkalinity regression. Type 1 background ranges were used preferentially. In the absence of data to calculate type 1 background ranges, type 2, followed by type 3 and type 4 background ranges were substituted. Further, background ranges were developed separately for each type of measurement available (dissolved, extractable, and total).

Relevant inorganic moieties for which background ranges were developed include aluminum, barium, beryllium, bismuth, copper, lithium, manganese, molybdenum, silver, tellurium, thallium, titanium, vanadium, and zinc. The corresponding PNECs chosen for the ERC-I were compared to the background ranges for each type of measurement (dissolved, extractable, and total), where available. The preliminary PNECs developed for these moieties were above the upper tolerance limits of the modelled background ranges in most ecozones (i.e., above reference levels).

#### **4.2.2 Predictive modelling: quantities in commerce**

Data on quantity of a substance in commerce (kg/yr) were gathered for all substances to which the ERC-I approach was applied. Quantity data consisted of chemical import or manufacture volume in Canada from recent section 71 surveys (Canada 2009; Canada 2012). Quantity data for most ERC-I substances came from results of Phase 2 of the DSL Inventory Update (Environment Canada 2013). In general, higher chemical quantities in commerce can be related to a higher probability of widespread exposure upon release to the environment.

Some substances not included in DSL Inventory Update surveys have relevant HS codes describing the substances as commodities, allowing import data to be collected to estimate quantities in commerce (CBSA 2016). Additionally, third-party market research reports were used to complement data from other sources and to fill gaps for substances not included in DSL Inventory Update surveys. Specifically, market research reports were obtained for hydrogen peroxide, lithium, and rare earth elements (CEH 2014a; CEH 2014b; CEH 2016; MRC 2016).

A near-field risk-based evaluation of all substances was performed to address the higher concentrations that may occur close to the point of discharge of a substance in the aquatic environment. In general, a conservative (precautionary) near-field scenario similar to that used in the rapid screening approach (EC, HC 2013; EC, HC 2014; ECCO, HC 2016) was employed to calculate PECs as described below.

The aquatic release scenario for near-field exposure involved applying a generic scenario to estimate local aquatic exposure. The equation and parameters used in this

scenario are given in Appendix B. While the generic aquatic exposure scenario has been developed to be conservative overall, the level of conservatism applied to individual parameters was selected to be moderate, since it is recognized that:

- a high level of conservatism applied to each parameter can easily compound into an excessively conservative overall exposure scenario;
- it is very unlikely that each parameter would be “worst case” at the same time; and
- interdependence of some parameters exists.

In brief, the scenario estimates exposure (PECs) in the aquatic environment using releases from an assumed, representative industrial facility that is manufacturing or using the substance. On the basis of the use codes and North American Industry Classification System (NAICS) codes provided in the DSL inventory update submissions, a generic emission factor of 2% (low), 25% (medium) or 100% (high) was assigned to the industrial activity. In order to do so, all use codes and NAICS were rated on the basis of professional judgement for their potential release to water. All undefined codes (U999) were rated manually after reviewing the description provided by the notifier. Assigned emission factors for each of the NAICS and use codes are independent of substance identity and are the same as those used for the ecological risk classification of organic substances (ECCC 2016a). Wastewater removal rates were by default the same value used in the rapid screening approach (EC, HC 2013; EC, HC 2014; ECCC, HC 2016), or they were replaced with median removal rates estimated from CMP Monitoring and Surveillance wastewater influent and effluent data (ECCC 2016b), where available, or with read-across information from other assessments. Refined removal rates were available for the inorganic moieties Bi, Ge, Mo, Sb, and V.

### 4.2.3 Predictive modelling: National Pollutant Release Inventory

Releases are reportable to the NPRI for a number of ERC-I substances and groups, as summarized in Table 4.4.

**Table 4-4. ERC-I substances and groups reportable to the NPRI**

<b>ERC-I substance or group</b>	<b>NPRI substance list name</b>	<b>NPRI CAS RN or other substance identifier</b>	<b>NPRI threshold category<sup>3</sup></b>	<b>ERC-I applicability</b>
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<sup>3</sup> NPRI Threshold Category Descriptions; Part 1A: A report is required for one or more substances if they were manufactured, processed or otherwise used (MPO) at a facility at a concentration of ≥1% by weight (except for by-products and mine tailings) and in a quantity of 10 tonnes or more, and employees worked 20,000 hours or more at a facility.



Antimony	Antimony (and its compounds)	NA - 01	Part 1A, 10 tonnes MPO	Conservative for 11 remaining priority substances
Lithium	Lithium carbonate	554-13-2	Part 1A, 10 tonnes MPO	Substance is a remaining priority
Molybdenum	Molybdenum trioxide	1313-27-5	Part 1A, 10 tonnes MPO	Substance is a remaining priority
Titanium	Titanium tetrachloride	7550-45-0	Part 1A, 10 tonnes MPO	Substance is a remaining priority
Vanadium	Vanadium (and its compounds)	7440-62-2	Part 1A, 10 tonnes MPO	Conservative for 2 remaining priority substances

A near-field risk-based evaluation was performed using NPRI data for reported releases to water for the last five years available in the downloadable dataset at the time of preparation (2011 to 2015). The calculations were similar to those performed on the data on quantities in commerce. However, instead of translating quantities in commerce to releases using estimated emission factors, the reported releases were used directly. The estimation of PECs from NPRI data is further detailed in Appendix B. High acute exposures in exceptional circumstances, such as large spills reported to the NPRI (e.g., tailings dam failures at the Mount Polley mine in 2014 and Obed Mountain coal mine in 2013) were excluded from PEC derivation for the purposes of determining an ecological risk classification for inorganics under typical conditions.

#### 4.2.4 Water quality monitoring

Water quality monitoring data for surface freshwaters were collected for each substance, where available, from the federal and provincial programs and repositories summarized in Table 3-1 for the period 2005 to 2015. Datasets from certain monitoring programs and repositories that were accessible online were queried and data were downloaded directly (EMS, JOSM, RAMP, PWQMN, and BQMA). Datasets from other sources were provided via direct request from program officials (SWQP, BEMLOSS, LTWQMN, SWQM, NLTWQM, and CMPEMS).

The federal and provincial programs and repositories included in the ERC-I approach were chosen in order to maximize coverage of the Canadian aquatic environment. These monitoring programs, including those captured by data repositories (e.g., EMS), may not have common mandates (e.g., to monitor waterbodies with shared political boundaries or to monitor waterbodies to gauge watershed health). While monitoring programs may serve different purposes, it appears that the majority of sampling locations were selected to monitor water quality under normal circumstances (i.e., not monitoring contamination from spills). Where available, data from programs monitoring

sectors (e.g., EEM from the MMER and wastewater effluent from the CMPEMS) were also included.

All datasets were processed individually in the following manner: field blanks, lab blanks, and suspected processing measurements (e.g., effluent measurements) were removed, and all units were converted to µg/L. Non-detects were substituted with half of the corresponding method detection limit (1/2 MDL). This is a standard but potentially over-conservative approach to addressing non-detects in monitoring data, and it still leaves uncertainty in the interpretation of non-detect data. This was considered in developing the profiling metrics and will be discussed further in the uncertainty section.

## **5. Preliminary classification**

Predictive modelling risk quotients were calculated for each theoretical facility from commercial quantity data (section 4.2.2) and for each facility reporting to the NPRI (section 4.2.3) by comparing calculated PECs with the PNECs derived in section 4.1.

Measured concentration (section 4.2.4) risk quotients were similarly calculated using the PNECs described in section 4.1. For PNECs that are functions of water quality parameters, paired water quality measurements were used when available, followed by station central tendencies (geomeans for [H+], hardness and dissolved organic carbon), then by provincial central tendencies. All data related to dissolved, extractable, and total measurements of the substances were included for analysis.

Through the calculated risk quotients, the hazard and exposure profiles for each inorganic substance were compared to decision criteria in order to assign an ecological risk classification. A qualitative score of low (L), moderate (M), or high (H) was given to each substance to represent lower, moderate, and higher levels of ecological risk potential from both predictive modelling and measured concentrations. These scores were then combined into a final classification. As previously indicated, the ERC-I science approach document only presents results with a final ecological risk classification of “low”. Ecological risks of “not-low” remaining priorities are subject to further evaluation of ecological considerations under the third phase of the CMP.

### **5.1 Risk classification criteria**

Ecological risk was classified by considering predictive modelling results from each data source and measured concentrations from each water quality monitoring dataset. The predictive modelling and measured concentrations were treated as two lines of evidence and combined in an overall risk classification ranking for each substance or group. Strength of each line of evidence varied for a given substance or group depending on data availability. Measured concentration scores were typically

qualitatively weighted higher than predictive modelling scores. The specific metrics are discussed below.

### 5.1.1 Predictive modelling classification

Ecological risk classifications for predictive modelling results were assigned to substances and groups of substances accounting for both the frequency and magnitude of PNEC exceedances (i.e., RQ greater than 1), for each data source (i.e., DSL-IU, NPRI, CBSA, and market research reports) according to the following criteria:

- High (H): More than one risk quotient (i.e., theoretical or real facility) > 10;
- Moderate (M): One risk quotient > 10 or More than one risk quotient between 1 and 10; or
- Low (L): At most one risk quotient between 1 and 10.

For the majority of substances, information generated from the DSL-IU1 or DSL-IU2 was the key input for predictive modelling. However, where NPRI release data, international trade data, or market research reports were also available for a given substance, multiple predictive modelling scores needed to be combined into a single classification. In determining the final predictive modelling classification, greater weight was assigned to modelling results using NPRI reported releases as input, followed by modelling results from DSL-IU survey data, international trade data, and market research.

**Table 5-1. Predictive modelling results for ERC-I substances and groups**

ERC-I substance or group (number of substances)	NPRI	DSL-IU	Import data	Market research	Overall predictive modelling classification
Antimony (11)	Low <sup>a</sup>	Moderate	Low <sup>f</sup>	NA	Low
Barium (4)	NA	Low	Low <sup>g</sup>	NA	Low
Beryllium (1)	NA	Moderate	NA	NA	Moderate
Bismuth (7)	NA	Low	NA	NA	Low
Cerium (1)	NA	Low	Moderate <sup>h</sup>	Low	Low
Deuterium oxide (1)	NA	Low	Low <sup>i</sup>	NA	Low
Germanium (1)	NA	Low	NA	NA	Low
Hydrogen peroxide (1)	NA	NA	Low <sup>j</sup>	Low	Low
Iodine (5)	NA	Low	Moderate <sup>k</sup>	NA	Low
Iron (2)	NA	Low	NA	NA	Low
Lanthanum (3)	NA	High	Moderate <sup>l</sup>	Low	Moderate
Lithium (16)	Low <sup>p</sup>	Low	Moderate <sup>m</sup>	Moderate	Low
Molybdenum (2)	Low <sup>c</sup>	Low	Low <sup>n</sup>	NA	Low
Neodymium (1)	NA	Low	Moderate <sup>l</sup>	Low	Low
Praseodymium (1)	NA	Low	Moderate <sup>l</sup>	Low	Low

Silicon carbide (1)	NA	NA	Low <sup>o</sup>	NA	Low
Sodium bromate (1)	NA	Low	NA	NA	Low
Talc (1)	NA	Low	Low <sup>p</sup>	NA	Low
Tellurium (2)	NA	Low	NA	NA	Low
Tin (2)	NA	NA	Low <sup>q</sup>	NA	Low
Titanium (13)	Low <sup>d</sup>	Low	Moderate <sup>r</sup>	NA	Low
Vanadium (2)	Low <sup>e</sup>	Low	Low <sup>s</sup>	NA	Low
Yttrium (1)	NA	Moderate	Moderate <sup>l</sup>	Low	Moderate

NA, not applicable; nes, not elsewhere specified

<sup>a</sup> Antimony (and its compounds)

<sup>b</sup> Lithium carbonate

<sup>c</sup> Molybdenum trioxide

<sup>d</sup> Titanium tetrachloride

<sup>e</sup> Vanadium (and its compounds)

<sup>f</sup> 2013 import data for antimony oxides (2825.80.0000) and 2011 import data for antimony potassium tartrates (2918.13.0010)

<sup>g</sup> 2013 import data for Barium sulphate (2833.27.0000), barium carbonate (2836.60.0000), barium chlorides (2827.39.0050 and 2827.39.0060) and 2011 import data for barium hydroxide (2816.40.0020)

<sup>h</sup> 2013 import data for Compounds, inorganic or organic, of rare-earth metals, of yttrium or of scandium or of mixtures of these metals, Cerium compounds (2846.10.0000)

<sup>i</sup> 2013 import data for Heavy water (deuterium oxide) (2845.10.0000)

<sup>j</sup> 2013 import data for Hydrogen peroxide: not solidified with urea (2847.00.0010)

<sup>k</sup> 2013 import data for Iodine (2801.20.0000), and 2011 import data for Manganous iodide; potassium iodide; sodium iodide (2827.60.1000)

<sup>l</sup> 2013 import data for Compounds, inorganic or organic, of rare-earth metals, of yttrium or of scandium or of mixtures of these metals, Other (2846.90.0000)

<sup>m</sup> 2013 import data for Lithium hydroxide (2825.20.0020), Lithium carbonates including pellets containing 95% or more by weight of lithium carbonates, for use in the manufacture of aluminum (2836.91.0010), and Lithium carbonates, other (2836.91.0090)

<sup>n</sup> 2013 import data for Molybdenum oxides and hydroxides (2825.70.0000)

<sup>o</sup> 2013 import data for Silicon carbide (2849.20.0000)

<sup>p</sup> 2013 import data for Natural steatite – Not crushed, not powdered (2526.10.0000), Natural steatite – Crushed or powdered: Talc of a particle size not exceeding 20 microns (2526.20.0010), and Natural steatite – Crushed or powdered: other (2526.20.0090)

<sup>q</sup> 2013 import data for Other articles of tin: Powders and flakes (8007.00.3000)

<sup>r</sup> 2013 import data for Titanium tetrachloride (2827.39.0030), Titanium dioxide (2823.00.0010), Titanium oxides, other (2823.00.0090), Pigment based on titanium dioxide, for use in Canadian manufactures (3206.11.1000), Pigments, containing 80% or more by weight of titanium dioxide, nes (3206.11.9010), Dispersions, containing 80% or more by weight of titanium dioxide, nes (3206.11.9020), Other preparations, based on titanium dioxide, nes (3206.19.9090)

<sup>s</sup> 2013 import data for Vanadium oxides and hydroxides (2825.30.0000)

## 5.1.2 Water quality monitoring classification

Ecological risk classifications for the water quality monitoring data were assigned to substance groups considering both the frequency and magnitude of PNEC exceedances (i.e., RQ greater than one) for each of the monitoring datasets, as identified in Table 3-1 (section 3), according to the criteria presented in Table 5-2.

**Table 5-2. Metrics for individual water quality monitoring datasets**

Metric	Low	Moderate	High
Frequency 1 (% of RQ>1)	0-5%	5–25%	> 25%

Frequency 2 (% of RQ>1 that are detected values)	0-5%	5–25%	> 25%
Magnitude 1 (median PNEC exceedance)	< PNEC (no exceedances)	1–10 x PNEC	> 10 x PNEC
Magnitude 2 (RQ of the 95 <sup>th</sup> percentile)	< PNEC	1–10 x PNEC	> 10 x PNEC

As indicated in Table 5-2, Frequency 1 is the percentage of PNEC exceedances (i.e., RQ greater than 1) within the dataset, Frequency 2 is the percentage of PNEC exceedances within the dataset corresponding to detected values only, Magnitude 1 is the median PNEC exceedance (if any), and Magnitude 2 is the risk quotient of the 95th percentile. Scores for these four metrics were combined to a single score for each monitoring dataset for all available substance groups, using the formula:

$$\text{Monitoring dataset score} = (\text{Number of metrics scored Low} \times 1) + (\text{Number of metrics scored Moderate} \times 10) + (\text{Number of metrics scored High} \times 100)$$

A classification for each individual monitoring dataset was then determined from the score according to Table 5-3.

**Table 5-3. Classifying individual monitoring datasets**

Monitoring dataset score	Monitoring dataset classification
< 40	Low
$40 \leq x < 200$	Moderate
$\geq 200$	High

The overall classification for each substance group was determined by the dataset with the highest classification. That is, if all monitoring datasets were classified as low, an overall classification of low was assigned. If one or more monitoring datasets were classified as moderate or high, an overall initial monitoring classification of moderate or high was initially assigned, respectively. However, substances or groups with initial water quality monitoring classifications of moderate or high were investigated further.

On the basis of these investigations, the following adjustments were made:

- Monitoring datasets where the extractable/total measurements scored moderate or high were adjusted to low if the corresponding dissolved measurements scored low. This adjustment was applied since the dissolved fraction is a closer representation of the bioavailable fraction and because the PNECs were derived from toxicological data for soluble test substances.
- In some instances, a single PNEC exceedance defining Magnitude 1, or multiple PNEC exceedances corresponding to less than three occurrences at different sampling sites captured by Magnitude 1 and/or Magnitude 2, resulted in monitoring dataset scores of moderate. Therefore, in these cases, the classifications were adjusted to low, given the exceptionally low Frequency 1 and/or Frequency 2 scores, and considering the conservative nature of

comparing single exceedances observed over periods of approximately 10 years to chronic PNEC values.

- One monitoring dataset for total bismuth, comprised solely of substituted non-detects (1/2 MDL), resulted in a high score for Frequency 1 and moderate scores for Magnitude 1 and Magnitude 2. However, given that the Frequency 2 score was low (i.e., zero), the classification was adjusted to low.
- The Saskatchewan (i.e., BEMLOSS) dataset for total lithium resulted in a moderate classification because of moderate scores for all metrics. The sampling sites with PNEC exceedances were mapped and showed no concordance with NPRI facilities releasing reportable lithium substances within a significant radius. There was also evidence that lithium concentrations in surface waters are elevated in the ecozone where the PNEC exceedances were observed (Prairies ecozone) due to geological factors (personal communication, email from the Water Science and Management Branch, Government of Manitoba, to the Ecological Assessment Division, Environment and Climate Change Canada, dated January 1, 2017; unreferenced). Background ranges were not available for lithium in Saskatchewan, which introduced uncertainty regarding the appropriateness of the PNEC for this ecozone. Therefore this dataset was reclassified to low.
- The Northwest Territories and BQMA datasets for some of the rare earth elements received a moderate classification and were also considered for manual adjustment. According to market research reports (CEH 2016), rare earth element production was not active in these regions at the time of collection of the monitoring data. Coupled with low confidence in the remaining priority rare earth element PNEC values, low confidence in the relationship between the remaining priority rare earth substances and water quality monitoring, and lack of data to establish background ranges in these regions, these were reclassified to low.

Further details regarding initial and adjusted classifications for the substance or groups resulting in an overall classification of low for water quality monitoring data are presented in the ERC-I supporting document (ECCC 2018).

## **6. Combined risk classification**

Risk classification scores from predictive modelling and monitoring dataset analyses were combined to arrive at a final ecological risk classification for inorganics. Following risk classification based on multiple criteria (sections 5.1.1 and 5.1.2), a risk matrix was used to classify the level of potential for ecological risk as high, moderate, or low. Table 6.1 lists the possible risk outcomes from combinations of predictive modelling and water quality monitoring classifications.

**Table 6-1. Risk matrix according to modelling and monitoring classifications**

	<b>Modelling low</b>	<b>Modelling moderate</b>	<b>Modelling high</b>
<b>Monitoring low</b>	Low	Low	Moderate
<b>Monitoring moderate</b>	Moderate	Moderate	Moderate
<b>Monitoring high</b>	Moderate	High	High

As reflected in Table 6.1, the water quality monitoring analysis was weighted higher than predictive modelling in determining the final risk classification score. Inorganic substances of higher risk concern would have higher magnitude and more frequent PNEC exceedances predicted or measured in the environment. Inorganic substances in the low ecological risk classification generally have smaller magnitude and infrequent PNEC exceedances, if any, whether measured or predicted.

## **7. Ecological risk classification results**

Appendix C lists 80 substances representing 12 chemical groups and 7 individual substances that were classified as being of overall low ecological concern. Additionally, 7 chemical groups were classified as being of overall high ecological concern (i.e., aluminum, copper, cyanide, manganese, silver, thallium, and zinc). Specific preliminary results from ERC-I for these 7 groups are not presented in this document as these have been scheduled for further consideration of ecological hazard and exposure in Type 3 assessments under the Chemicals Management Plan (Canada 2016; Canada 2017). For the 80 substances classified as being of low ecological concern, the results presented in this Science Approach Document will form the basis, in conjunction with any other relevant information that becomes available after its publication, of conclusions of screening assessments to be published under section 68 or 74 of CEPA.

The ERC-I classification result is intended to apply to specific CAS RNs identified as remaining priorities and should not be interpreted as applicable to corresponding moieties. For example, this is particularly true in the case of the “rare earth elements,” where 7 substances covering only 5 elements would be a particularly poor representation of this chemical class.

Additionally, substances classified as low ecological concern with relatively high hazard (e.g., PNECs in low parts per billion) but which are currently reported to be produced, imported, or used in low volume in Canada (e.g., rare earth elements, bismuth), are proposed to be identified for additional tracking of production and use patterns. Their priority status may be re-evaluated if new information becomes available.

**Table 7-1. Percentage breakdown of the final ecological risk classification for inorganic substances**

<b>Ecological risk classification</b>	<b>Substance groups and individuals</b>	<b>Percent (%)</b>
Low (use pattern data is proposed to be collected for some of these moieties)	19	73
Moderate or high (to be included in future group for refined assessment of ecological risk)	7	27
<b>Total</b>	<b>26</b>	<b>100</b>

Critical data and considerations used to create substance-specific profiles and classifications associated with hazard, exposure and risk are presented in ECCC 2018.

## **8. Assessment of risk classification uncertainty**

The ERC-I generally reflects a conservative or precautionary approach where multiple metrics add to a weight of evidence for classification. Given that the main purpose of ERC-I was to identify substances and groups of potentially lower ecological concern, the approach generally sought to minimize the potential for under-classification of ecological risk.

A simplifying assumption taken for ERC-I was to focus on the aquatic compartment, which left uncertainty as to hazard and exposure in air, soil, and sediment. Data availability was a consideration in focussing on the aquatic compartment for ERC-I, the freshwater aquatic compartment being the most likely to have a basic set of hazard information to compare across substances, due to a historical preference for testing aquatic species for ecological effects (Swanson and Socha 1997). Also, as the ecological remaining priorities were identified as such on the basis of water solubility exceeding an aquatic toxicity threshold (Canada 2007), the potential for false negatives in the classification of ecological priority substances as being of low ecological concern based on the aquatic compartment is believed to be minimal.

Hazard profiling involved the review for applicability and alignment or derivation of chronic aquatic PNECs for inorganic substances and groups. Conservative assumptions made in the near-field exposure modelling (e.g., the use of generic emissions rates) reduced the potential for under-classifying ecological risk. The use of water quality monitoring data collected over approximately 10 years increased the confidence in the datasets. This led to the attribution of higher weight to the water quality monitoring dataset analyses in determining the final ecological risk classification. Another consideration would be the wide geographical and seasonal diversity captured in the



water quality monitoring datasets, compared to the fixed conservative values of the near-field exposure scenario.

Exposure profiling included the use of commercial data (quantity and use) from a number of sources. Information from the DSL-IU was inherently specific to the individual remaining priority substances that were surveyed. However, international trade data is collected on the basis of HS Codes, which introduced uncertainty in relating the information to remaining priority CAS RNs. HS Codes may encompass multiple CAS RNs for related substances, not all of which are remaining priorities. For example, the HS Code 2827.60.1000 has the 2011 consolidated description “manganous iodide; potassium iodide; sodium iodide,” but only potassium iodide and sodium iodide are remaining priority substances. In these situations, relating the import data from the broader HS Code to the priority substances reduces the possibility of under-classifying risk.

Similarly, the water quality monitoring data considered in ERC-I included dissolved, extractable, and total measurements, reflecting multiple exposures to moieties of potential concern. This introduced an inconsistency when applying results from the water quality monitoring datasets, which are moiety-based, to the subset of remaining priority substances. However, these comparisons are conservative from the perspective of the remaining priority substances, which further reduces the possibility of under-classifying risk. Non-detect values in the water quality monitoring datasets were substituted with half the corresponding detection limit (1/2 MDL). This is a standard but potentially over-conservative approach to addressing non-detects in monitoring data. The Frequency 2 metric was introduced to reduce the weight assigned to non-detect measurements in exceedance of PNECs. Finally, water quality monitoring data were not available from every data source for every substance in ERC-I. This may introduce some bias toward substances which are measured and reported more frequently. However, it was deemed appropriate to use all identifiable data rather than limit the approach to substances with rich monitoring datasets. In addition, the IRAP provides a mechanism to revisit currently data-poor substances if data becomes available.

## **9. Conclusion**

On the basis of inherent hazard properties, current use patterns, quantities in commerce, available release data, and an analysis of federal and provincial water quality monitoring data, 80 substances were classified as being of low ecological concern. Taking into account current information, the 80 substances of low ecological concern are not expected to pose an ecological risk. The approach and results for these 80 substances will form the basis, in conjunction with any other relevant information that becomes available after the publication of this Science Approach Document, for the conclusions in screening assessments that will be published at a later time. Further follow-up or tracking of information may be done for substances that were classified as

low concern primarily on the basis of current low exposures, to determine whether further activity is required in the future.

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## Appendices

### Appendix A. Exposure data availability for substances classified as being of low ecological concern by the ERC-I approach

CMP chemical group <sup>a</sup>	CAS RN	Domestic Substances List name	DSL-IU1 (2008)	DSL-IU2 (2011)	NPRI	CBSA <sup>b</sup>	WQM (number of datasets) <sup>c</sup>
Antimony	1314-60-9	Antimony oxide (Sb <sub>2</sub> O <sub>5</sub> )	No	Yes	Yes <sup>d</sup>	Yes	9
Antimony	1327-33-9	Antimony oxide	Yes	No	Yes <sup>d</sup>	Yes	9
Antimony	1345-04-6	Antimony sulfide (Sb <sub>2</sub> S <sub>3</sub> )	Yes	No	Yes <sup>d</sup>	No	9
Antimony	10025-91-9	Stibine, trichloro-	Yes	No	Yes <sup>d</sup>	No	9
Antimony	15432-85-6	Antimonate (SbO <sub>3</sub> <sup>1-</sup> ), sodium	Yes	No	Yes <sup>d</sup>	No	9
Antimony	15874-48-3	Phosphorodithioic acid, O,O-dipropyl ester, antimony(3+)salt	Yes	No	Yes <sup>d</sup>	No	9
Antimony	15890-25-2	Antimony, tris(dipentylcarbamodithioato-S,S)-, (OC-6-11)-	Yes	No	Yes <sup>d</sup>	No	9
Antimony	15991-76-1	Antimony, tris[bis(2-ethylhexyl)carbamodithioato-S,S]-, (OC-6-11)-	Yes	No	Yes <sup>d</sup>	No	9
Antimony	28300-74-5	Antimonate(2-), bis[μ-[2,3-di(hydroxy-κO)butanedioato(4-)-κO <sup>1</sup> :κO <sup>4</sup> ]]di-, dipotassium, trihydrate, stereoisomer	Yes	No	Yes <sup>d</sup>	Yes	9
Antimony	29638-69-5	Antimonate (Sb <sub>2</sub> O <sub>7</sub> <sup>4-</sup> ), tetrapotassium	Yes	No	Yes <sup>d</sup>	No	9
Antimony	33908-66-6	Antimonate (Sb(OH) <sub>6</sub> <sup>1-</sup> ), sodium, (OC-6-11)-	Yes	No	Yes <sup>d</sup>	No	9
Barium	513-77-9	Carbonic acid, barium salt (1:1)	No	No	No	Yes	10
Barium	7727-43-7	Sulfuric acid, barium salt (1:1)	No	No	No	Yes	10
Barium	10361-37-2	Barium chloride (BaCl <sub>2</sub> )	No	Yes	No	Yes	10
Barium	17194-00-2	Barium hydroxide (Ba(OH) <sub>2</sub> )	No	Yes	No	Yes	10
Beryllium	7440-41-7	Beryllium	Yes	No	No	No	9

<b>CMP chemical group<sup>a</sup></b>	<b>CAS RN</b>	<b>Domestic Substances List name</b>	<b>DSL-IU1 (2008)</b>	<b>DSL-IU2 (2011)</b>	<b>NPRI</b>	<b>CBSA<sup>b</sup></b>	<b>WQM (number of datasets)<sup>c</sup></b>
Bismuth	1304-76-3	Bismuth oxide	No	Yes	No	No	9
Bismuth	1304-85-4	Bismuth hydroxide nitrate oxide (Bi <sub>5</sub> (OH) <sub>9</sub> (NO <sub>3</sub> ) <sub>4</sub> O)	No	Yes	No	No	9
Bismuth	10361-44-1	Nitric acid, bismuth(3+) salt	No	Yes	No	No	9
Bismuth	14059-33-7	Bismuth vanadium oxide (BiVO <sub>4</sub> )	Yes	No	No	No	9
Bismuth	21260-46-8	Bismuth, tris(dimethylcarbamodithioato-S,S')-, (OC-6-11)-	No	Yes	No	No	9
Bismuth	34364-26-6	Neodecanoic acid, bismuth(3+) salt	No	Yes	No	No	9
Bismuth	67874-71-9	Hexanoic acid, 2-ethyl-, bismuth(3+) salt	No	Yes	No	No	9
Bromate	7789-38-0	Bromic acid, sodium salt	Yes	No	No	No	NA
Individual	409-21-2	Silicon carbide (SiC)	No	No	No	Yes	NA
Individual	7722-84-1	Hydrogen peroxide (H <sub>2</sub> O <sub>2</sub> )	No	No	No	Yes	NA
Individual	7789-20-0	Water-d2	No	Yes	No	Yes	NA
Individual	10038-98-9	Germane, tetrachloro-	No	Yes	No	No	NA
Individual	14807-96-6	Talc (Mg <sub>3</sub> H <sub>2</sub> (SiO <sub>3</sub> ) <sub>4</sub> )	No	Yes	No	Yes	NA
Iodine	7553-56-2	Iodine	No	Yes	No	Yes	NA
Iodine	7681-11-0	Potassium iodide (KI)	No	Yes	No	Yes	NA
Iodine	7681-82-5	Sodium iodide (NaI)	No	Yes	No	Yes	NA
Iodine	20461-54-5	Iodide	No	Yes	No	No	NA
Iodine	63325-16-6	Mercury, diiodobis(5-iodo-2-pyridinamine)-, dihydriodide	No	No	No	No	NA
Iron	12713-03-0	Umber	No	Yes	No	No	NA <sup>f</sup>
Iron	51274-00-1	C.I. Pigment Yellow 42	No	Yes	No	No	NA <sup>f</sup>
Lithium	546-89-4	Acetic acid, lithium salt	No	Yes	No	No	9
Lithium	554-13-2	Carbonic acid, dilithium salt	No	Yes	Yes	Yes	9
Lithium	1310-65-2	Lithium hydroxide (Li(OH))	No	Yes	No	Yes	9
Lithium	4485-12-5	Octadecanoic acid,	No	Yes	No	No	9

<b>CMP chemical group<sup>a</sup></b>	<b>CAS RN</b>	<b>Domestic Substances List name</b>	<b>DSL-IU1 (2008)</b>	<b>DSL-IU2 (2011)</b>	<b>NPRI</b>	<b>CBSA<sup>b</sup></b>	<b>WQM (number of datasets)<sup>c</sup></b>
		lithium salt					
Lithium	7439-93-2	Lithium	No	Yes	No	No	9
Lithium	7447-41-8	Lithium chloride (LiCl)	No	Yes	No	No	9
Lithium	7620-77-1	Octadecanoic acid, 12-hydroxy-, monolithium salt	No	Yes	No	No	9
Lithium	7789-24-4	Lithium fluoride (LiF)	No	Yes	No	No	9
Lithium	10377-48-7	Sulfuric acid, dilithium salt	No	Yes	No	No	9
Lithium	12627-14-4	Silicic acid, lithium salt	No	Yes	No	No	9
Lithium	13840-33-0	Hypochlorous acid, lithium salt	No	Yes	No	No	9
Lithium	27253-30-1	Neodecanoic acid, lithium salt	No	Yes	No	No	9
Lithium	38900-29-7	Nonanedioic acid, dilithium salt	No	Yes	No	No	9
Lithium	53320-86-8	Silicic acid, lithium magnesium sodium salt	No	Yes	No	No	9
Lithium	68783-37-9	Fatty acids, C16-18, lithium salts	No	Yes	No	No	9
Lithium	68649-48-9	Paraffin waxes and Hydrocarbon waxes, oxidized, lithium salts	No	No	No	No	9
Molybdenum	1313-27-5	Molybdenum oxide (MoO <sub>3</sub> )	Yes	No	Yes	Yes	9
Molybdenum	1317-33-5	Molybdenum sulfide (MoS <sub>2</sub> )	No	Yes	No	No	9
Rare Earth Elements	1312-81-8	Lanthanum oxide (La <sub>2</sub> O <sub>3</sub> )	No	Yes	No	Yes	4
Rare Earth Elements	1314-36-9	Yttrium oxide (Y <sub>2</sub> O <sub>3</sub> )	No	Yes	No	Yes	2
Rare Earth Elements	10099-58-8	Lanthanum chloride (LaCl <sub>3</sub> )	No	Yes	No	Yes	4
Rare Earth Elements	12008-21-8	Lanthanum boride, (OC-6-11)-	No	Yes	No	Yes	4
Rare Earth Elements	12036-32-7	Praseodymium oxide (Pr <sub>2</sub> O <sub>3</sub> )	No	Yes	No	Yes	1
Rare Earth Elements	56797-01-4	Hexanoic acid, 2-ethyl-, cerium(3+) salt	No	Yes	No	Yes	4





<sup>b</sup> Harmonized System (HS) codes are not equivalent to CAS RNs. HS code descriptions frequently cover multiple CAS RNs. An entry of "yes" indicates the CAS RN has a clear relationship to an HS description, as identified in the Customs Tariff Departmental Consolidation.

<sup>c</sup> Water quality monitoring performed on a total, dissolved, or extractable inorganic basis, not specific to a CAS RN.

<sup>d</sup> Antimony (and its compounds) is on the List of NPRI Substances (Total of the pure element and the equivalent weight of the element contained in any compound, alloy or mixture).

<sup>e</sup> Vanadium (and its compounds) is on the List of NPRI Substances (Total of pure vanadium and the equivalent weight of vanadium contained in any compound or mixture. Excludes vanadium contained in an alloy).

<sup>f</sup> Water quality monitoring data available is available for iron, but was deemed inappropriate for comparison to two insoluble pigments that are the remaining priorities.

## Appendix B. Summary of local exposure screening scenarios

- A conservative predicted environmental concentration (PEC) resulting from the release of the substance to the aquatic environment from an industrial point source is calculated as shown in the following equation. Parameters used in this exposure scenario are described in Table B-1 below. Default values indicated in Table B-1 were replaced with substance- and notifier-specific information when available.

$$\text{PEC (mg/L)} = (\text{Qty} \times \text{Metal content} \times \text{Release} \times (1 - \text{Wastewater removal})) / (\text{Duration} \times (\text{River flow} + \text{Wastewater flow})) \times (1000/86400)$$

- For National Pollutant Release Inventory (NPRI) data, the numerator of the above equation is replaced with the annual quantity reported to be released. For substances not surveyed in a DSL Inventory Update, quantities were obtained from international trade data and market research reports. Emission factors to pair with quantities from international trade data and market research reports were assigned on the basis of professional judgement. For one substance with particularly high consumer use quantities, a consumer release aquatic model was used to generate PECs as an additional line of evidence.
- The PEC is compared to the PNEC (determined as per Section 4.1) to calculate a risk quotient (PEC / PNEC). If the risk quotient is greater than 1, this indicates that the conservatively estimated concentration in water exceeds the aquatic estimated no-effect level and that there exists a potential to cause harm in the aquatic ecosystem. A value below 1 indicates that concentrations that may cause an effect to sensitive aquatic organisms are not reached and therefore harm to aquatic organisms is less likely under this scenario.
- A risk quotient was determined for each theoretical facility, representing one of the following: a notifier to the DSL Inventory Update (phases one or two), an importer identified in international trade data, or a company identified in market research reports, associated with the substances of interest. Risk quotients were also determined from NPRI release data.

**Table B-1. Parameters used in near-field scenario**

Abbreviation	Parameter	Value	Units	Notes
Qty	Quantity of substance reported by notifier	Quantity from Inventory Update, CBSA, or market research	kg/year	Substance-specific

Metal Content	Weight percentage of inorganic moiety present in a substance	Substance specific	%	Determined from available substance identity information, conservative assumption of 100% when no information available
Release	Release of substance during industrial process	2% (low) 25% (medium) 100% (high)	%	Default value, independent of substance identity, based on analysis of reported use and NAICS codes (ECCC 2016a)
Wastewater Removal	Wastewater treatment system (WWTS) removal efficiency	Default value = 70 (SimpleTreat model not applicable to inorganics)	%	Default value replaced with substance-specific information when available (ECCC 2016b)
Duration	Duration over which substance is released	150	days/ year	Assumes variable or discontinuous use of substance over a year
Wastewater Flow	WWTS flow rate	0.04	m <sup>3</sup> /s	10th percentile of municipal WWTS flow rates in Canada
River Flow	Flow of receiving watercourse	1.84	m <sup>3</sup> /s	15th percentile of the distribution of receiving watercourse flows in the country (based on the distribution of the 50th percentile of flow rates); weighted by number of industries releasing to the receiving watercourse
Not applicable	Factor combining conversion from kg to mg and m <sup>3</sup> to L	1000	mg·m <sup>3</sup> / kg·L	Not applicable
Not applicable	Conversion factor from seconds to days	86400	seconds /day	Not applicable

## Appendix C. Substances classified as being of low ecological concern

Table C-1. Substances classified as being of low ecological concern

CMP chemical group <sup>a</sup>	CAS RN	Domestic Substances List name	Predictive modelling ranking	Water quality monitoring ranking	Overall ERC-I ranking
Antimony	1314-60-9	Antimony oxide (Sb <sub>2</sub> O <sub>5</sub> )	Low	Low	Low
Antimony	1327-33-9	Antimony oxide	Low	Low	Low
Antimony	1345-04-6	Antimony sulfide (Sb <sub>2</sub> S <sub>3</sub> )	Low	Low	Low
Antimony	10025-91-9	Stibine, trichloro-	Low	Low	Low
Antimony	15432-85-6	Antimonate (SbO <sub>3</sub> <sup>1-</sup> ), sodium	Low	Low	Low
Antimony	15874-48-3	Phosphorodithioic acid, O,O-dipropyl ester, antimony(3+)salt	Low	Low	Low
Antimony	15890-25-2	Antimony, tris(dipentylcarbamodithioat o-S,S')-, (OC-6-11)-	Low	Low	Low
Antimony	15991-76-1	Antimony, tris[bis(2-ethylhexyl)carbamodithioat o-S,S']-, (OC-6-11)-	Low	Low	Low
Antimony	28300-74-5	Antimonate(2-), bis[μ-[2,3-di(hydroxy-κO)butanedioato(4-)-κO1:κO4]]di-, dipotassium, trihydrate, stereoisomer	Low	Low	Low
Antimony	29638-69-5	Antimonate (Sb <sub>2</sub> O <sub>7</sub> <sup>4-</sup> ), tetrapotassium	Low	Low	Low
Antimony	33908-66-6	Antimonate (Sb(OH) <sub>6</sub> <sup>1-</sup> ), sodium, (OC-6-11)-	Low	Low	Low
Barium	513-77-9	Carbonic acid, barium salt (1:1)	Low	Low	Low
Barium	7727-43-7	Sulfuric acid, barium salt (1:1)	Low	Low	Low
Barium	10361-37-2	Barium chloride (BaCl <sub>2</sub> )	Low	Low	Low
Barium	17194-00-2	Barium hydroxide (Ba(OH) <sub>2</sub> )	Low	Low	Low
Beryllium	7440-41-7	Beryllium	Moderate	Low	Low
Bismuth	1304-76-3	Bismuth oxide	Low	Low	Low
Bismuth	1304-85-4	Bismuth hydroxide nitrate oxide (Bi <sub>5</sub> (OH) <sub>9</sub> (NO <sub>3</sub> ) <sub>4</sub> O)	Low	Low	Low
Bismuth	10361-44-1	Nitric acid, bismuth(3+) salt	Low	Low	Low

<b>CMP chemical group<sup>a</sup></b>	<b>CAS RN</b>	<b>Domestic Substances List name</b>	<b>Predictive modelling ranking</b>	<b>Water quality monitoring ranking</b>	<b>Overall ERC-I ranking</b>
Bismuth	14059-33-7	Bismuth vanadium oxide (BiVO <sub>4</sub> )	Low	Low	Low
Bismuth	21260-46-8	Bismuth, tris(dimethylcarbamodithioato-S,S'), (OC-6-11)-	Low	Low	Low
Bismuth	34364-26-6	Neodecanoic acid, bismuth(3+) salt	Low	Low	Low
Bismuth	67874-71-9	Hexanoic acid, 2-ethyl-, bismuth(3+) salt	Low	Low	Low
Bromate	7789-38-0	Bromic acid, sodium salt	Low	NA	Low
Individual	409-21-2	Silicon carbide (SiC)	Low	NA	Low
Individual	7722-84-1	Hydrogen peroxide (H <sub>2</sub> O <sub>2</sub> )	Low	NA	Low
Individual	7789-20-0	Water-d2	Low	NA	Low
Individual	10038-98-9	Germane, tetrachloro-	Low	NA	Low
Individual	14807-96-6	Talc (Mg <sub>3</sub> H <sub>2</sub> (SiO <sub>3</sub> ) <sub>4</sub> )	Low	NA	Low
Iodine	7553-56-2	Iodine	Low	NA	Low
Iodine	7681-11-0	Potassium iodide (KI)	Low	NA	Low
Iodine	7681-82-5	Sodium iodide (NaI)	Low	NA	Low
Iodine	20461-54-5	Iodine	Low	NA	Low
Iodine	63325-16-6	Mercury, diiodobis(5-iodo-2-pyridinamine)-, dihydriodide	Low	NA	Low
Iron	12713-03-0	Umber	Low	NA	Low
Iron	51274-00-1	C.I. Pigment Yellow 42	Low	NA	Low
Lithium	546-89-4	Acetic acid, lithium salt	Low	Low	Low
Lithium	554-13-2	Carbonic acid, dilithium salt	Low	Low	Low
Lithium	1310-65-2	Lithium hydroxide (Li(OH))	Low	Low	Low
Lithium	4485-12-5	Octadecanoic acid, lithium salt	Low	Low	Low
Lithium	7439-93-2	Lithium	Low	Low	Low
Lithium	7447-41-8	Lithium chloride (LiCl)	Low	Low	Low
Lithium	7620-77-1	Octadecanoic acid, 12-hydroxy-, monolithium salt	Low	Low	Low
Lithium	7789-24-4	Lithium fluoride (LiF)	Low	Low	Low
Lithium	10377-48-7	Sulfuric acid, dilithium salt	Low	Low	Low
Lithium	12627-14-4	Silicic acid, lithium salt	Low	Low	Low



<b>CMP chemical group<sup>a</sup></b>	<b>CAS RN</b>	<b>Domestic Substances List name</b>	<b>Predictive modelling ranking</b>	<b>Water quality monitoring ranking</b>	<b>Overall ERC-I ranking</b>
		salt			
Titanium	1070-10-6	1-Hexanol, 2-ethyl-, titanium(4+) salt	Low	Low	Low
Titanium	1317-80-2	Rutile	Low	Low	Low
Titanium	1344-54-3	Titanium oxide	Low	Low	Low
Titanium	5593-70-4	1-Butanol, titanium(4+) salt	Low	Low	Low
Titanium	7550-45-0	Titanium chloride (TiCl <sub>4</sub> ) (T-4)-	Low	Low	Low
Titanium	7705-07-9	Titanium chloride (TiCl <sub>3</sub> )	Low	Low	Low
Titanium	12047-27-7	Titanate, barium (1:1)	Low	Low	Low
Titanium	12060-59-2	Titanate, strontium (1:1)	Low	Low	Low
Titanium	13463-67-7	Titanium oxide (TiO <sub>2</sub> )	Low	Low	Low
Titanium	13825-74-6	Titanium, oxo[sulfato(2-)-kO,kO']-	Low	Low	Low
Titanium	16919-27-0	Titanate(2-), hexafluoro-, dipotassium, (OC-6-11)-	Low	Low	Low
Titanium	20338-08-3	Titanium hydroxide (Ti(OH) <sub>4</sub> ) (T-4)-	Low	Low	Low
Vanadium	7727-18-6	Vanadium, trichlorooxo-	Low	Low	Low
Vanadium	11099-11-9	Vanadium oxide	Low	Low	Low

Abbreviations: CAS RN, Chemical Abstracts Service Registry Number; CMP, Chemicals Management Plan; ERC-I, Ecological Risk Classification of Inorganic Substances; NA, not available

<sup>a</sup> Note that the CMP chemical group assignment only serves for data organization and conservative comparisons to water quality monitoring data where available.